

Comparative Toxicity of Endosulfan and Fipronil Insecticides: Utilizing *In Vivo* and *In Vitro* Data

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Abstract

Background: Endosulfan, an organochlorine compound, and fipronil, a phenylpyrazole, are insecticides with a common mechanism of toxicity. They interfere with Cl^- influx by binding to the gamma-aminobutyric acid receptor (GABA_AR) and blocking the inhibitory actions of GABA_A . *In vivo* they cause neurotoxicity, hepatotoxicity, developmental toxicity, and can alter endocrine and immune systems. The thyroid is a target of fipronil toxicity. Human exposure occurs via food residues, skin contact and/or air dispersion. They are environmentally persistent and bioaccumulate in food chains.

Method: Compare *in vivo* data with *in vitro* results from the U.S. Environmental Protection Agency (USEPA) Toxicology Forecaster (ToxCast) high-throughput screening assays and zebrafish models to assess their usefulness in predicting toxicity.

Results: Fipronil's *in vivo* toxicity occurred at lower doses than endosulfan for similar effects. ToxCast was a weak predictor of liver toxicity and estrogen receptor interaction. Missing is evidence of "true actives" for fipronil ToxCast assays with the thyroid receptor and for either compound with GABA_AR or androgen receptors. Zebrafish models were good predictors of endosulfan and fipronil neurotoxicity in mammalian *in vivo* studies.

Conclusion: ToxCast assays do not provide support for *in vivo* neurotoxicity or endocrine disruption where zebrafish are good predictors of both parameters.

Key Words: Endosulfan; Fipronil; Organochlorine; GABA Inhibitors; Toxcast; Zebrafish; Carcinogens

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Introduction

Endosulfan (ES: 6, 7, 8, 9, 10, 10-hexachloro-1, 5, 5 α , 6, 9, 9 α -hexahydro-6, 9-methano-2, 4, 3-benzodioxathiepin-3-oxide; [isomers: α -; β -]) and fipronil (FP: (RS)-5-amino-1-[2,6-dichloro-4-(trifluoromethyl)phenyl]-4-(trifluoromethylsulfinyl)-1H-pyrazole-3-carbonitrile) represent the first and second generations, respectively, of chloride channel blocker insecticides whose primary target is the central nervous system [1, 2]. High potential for human exposure is indicated for both pesticides. Exposure to general public results from ingesting residues in food, inhaling vapors, skin contact. Workers' exposures are mainly dermal and inhalation from handling or re-entering treated fields. A large number of poisonings are reported from fipronil's use in

veterinary products for flea and tick control. The exposure is likely higher for general public living near hazardous waste sites where endosulfan has been detected [3, 4]. As noncompetitive antagonists of the GABA_A -receptor, they block passage of Cl^- ions and thereby the actions of the inhibitory neurotransmitter GABA [2]. In mammals, seizures, vomiting and convulsions are symptoms of excessive nerve stimulation associated with GABA_A antagonism [5]. Their action in the brain is very complex due to biotransformation to toxic metabolites that can act on GABA_AR (e.g., FP sulfone [6]). Each is detoxified by P450s (CYP2B6, CYP3A4-5: ES; hCYP2C9: FP) [7, 8] but neither requires activation to be toxic⁸. ES and FP are associated with endocrine disruption. FP has induced thyroid and liver cancer in animal models (<http://www.cdpr.ca.gov/>

docs/risk/toxsums/toxsumlist.htm; accessed 9/2015) [9]. These well characterized *in vivo* effects provide a basis for comparing toxicities.

Another means of comparison is the U.S. Environmental Protection Agency (USEPA) Toxicology Forecaster (ToxCast) *in vitro* high-throughput screening program; profiling more than 1,000 chemicals (~ 800 assays <http://actor.epa.gov/dashboard2/>; accessed 9/2015). In addition, zebrafish models are a rapid *in vivo* method for assessing developmental effects [10-13]. The goal of this brief communication is to examine if ToxCast and zebrafish results were predictive of targets and activities relevant to *in vivo* ES and FP endpoints.

Method

In vivo studies

In vivo studies in **Table 1** summarize relevant neurotoxic or endocrine disrupting effects for both chemicals. No-Observed-Effect-Levels (NOEL) and Lowest-Observed-Effect-Levels (LOEL) were established by the California Department of Pesticide Regulation from guideline studies required for pesticide registration (<http://www.cdpr.ca.gov/docs/risk/rcd.htm>) and from open literature.

In vitro studies

We accessed the recently updated ToxCast database (<http://actor.epa.gov/dashboard2/>) for active assays that could inform effects manifested in overt toxicity in *in vivo* studies.

Zebrafish studies

Data from two published methods included embryos treated with chorion intact [10, 11] or with chorion removed [12, 13].

In vivo to *in vitro* extrapolation (IVIVE)

IVIVE was used to convert zebrafish NOELs expressed as μM

concentrations into oral equivalent doses (OED: mg/kg/d). Estimated pharmacokinetic data derived from linear regression, published *in vitro* hepatic metabolism and protein binding with **rat data** are used for the conversion [14].

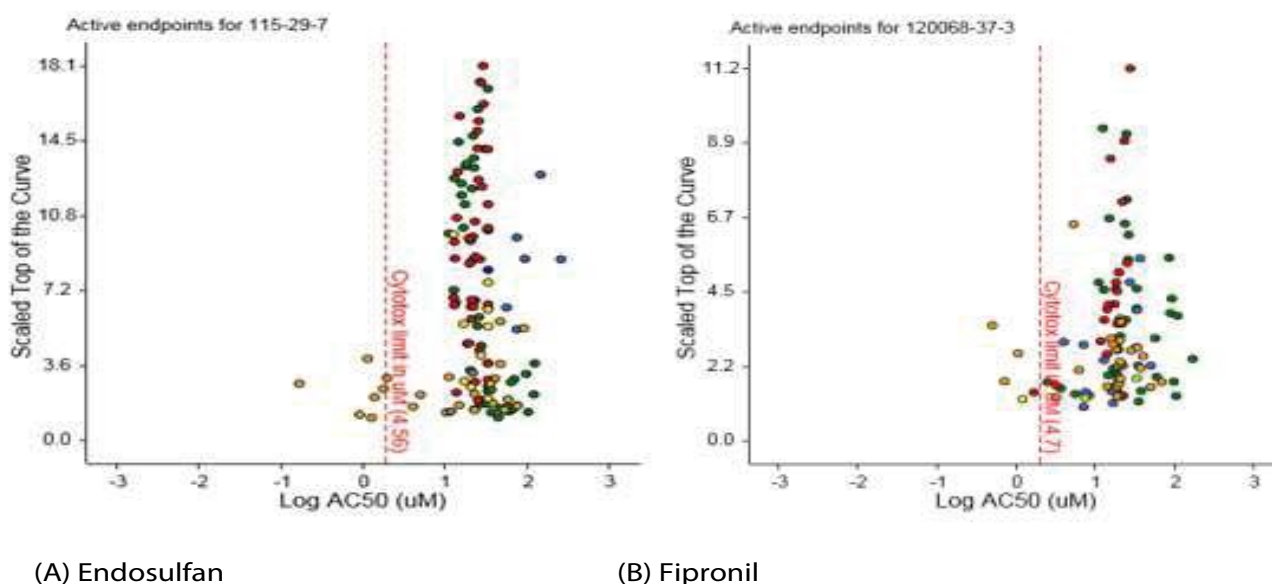
Results

In vivo

Both insecticides were neurotoxic [15, 16] with LOELs of 2-2.5 mg/kg/d in rats and dogs (**Table 1**). ES developmental effects in rat pups and fetuses occurred at up to 7-fold higher doses than FP. Potential endocrine disruption was shown in both compounds by developmental delays and skeletal variations. Liver effects in rats and mice at LOELs of 3.95 mg/kg/d (ES) and 0.13 mg/kg/d (FP) included increased liver weights and hepatocellular carcinomas in mice (FP [17]). FP thyroid toxicity was evident by altered thyroid function and follicular cell adenomas in rats (LOEL 0.06 mg/kg/d) [18].

ToxCast data

Results are reported as AC_{50} s ($\frac{1}{2}$ maximal activity concentrations: Dashboard: <http://actor.epa.gov/dashboard/2>). All active assays appear in **Figure 1** (dots) to the right and left of the cytotoxicity limit; however true actives have AC_{50} s below the cytotoxicity limits (e.g. $\leq 4.56 \mu\text{M}$ ES; $\leq 4.7 \mu\text{M}$ FP). Active assays beyond the cytotoxicity limit may not be specific to a chemical-receptor interaction but represent a burst of cellular responses indicative of cytotoxicity. ES true actives were assays having to do with xenobiotic metabolism (human constitutive androstane receptor: hCAR; immune response: DR5; pregnane-x: affecting hCYP2B6 genes; hCYP3A4 xenobiotic response element: PXRE) and the estrogen receptor ($ER\alpha$; estrogen response element: ERE). True actives for FP were related to cell adhesion molecules (kinases), hCYP2C9, xenobiotic metabolism (hCAR) and ERE. **Figure 2** shows



The red line indicates the region where cytotoxicity begins. True actives are to the left of that line; to the right cytotoxicity increases with increasing Log AC_{50} s. AC_{50} = activity at $\frac{1}{2}$ maximal concentration.

Figure 1 ES (A) and FP (B) active ToxCast assays.

active versus inactive assays and associated components for each compound (abbreviations at: <http://www.epa.gov/comptox/toxcast/data.html>).

Zebrafish data

Zebrafish had an AC_{50} of 1 μ M after exposure to ES (chorion intact) [10]; with a Toxicity Score (all malformations) of 40 (highest possible score) at peak concentrations (4 μ M). When embryo chorions were removed, ES was highly toxic to larval development at ≥ 0.0064 μ M [12]. ES α -treated embryos (chorion intact) showed neurobehavioral effects (abnormal swim behavior, disorientation, abnormal touch response) at ≥ 1.0 μ M (NOEL 0.5 μ M) [22].

FP-treated zebrafish (chorion intact) had an AC_{50} of 15.5 μ M with a "Toxicity Score" of 40 [10]. Without the chorion, however, FP had no effects on development [10]. Zebrafish (chorion removed) treated with FP at ≥ 0.23 μ M had irreversible effects on body length, notochord degeneration, abnormal axial muscle morphology, muscle fiber degeneration [11].

IVIVE for ES neurotoxicity in zebrafish (chorion removed) [13] estimated an OED of 0.019 mg/kg/d (0.5 μ M NOEL) [9]; 25-fold lower than the neurotoxicity NOEL in dogs (0.5 mg/kg/d) (Table 1) [15]. For FP the zebrafish OED was 0.37 mg/kg/d based on neurotoxicity at 23 μ M [13]. This OED is equivalent to the *in vivo* NOEL (0.2 mg/kg/d) from an acute neurotoxicity rat study [16].

Discussion/Conclusion

Data indicate that *in vivo*, FP's toxicity occurred at lower doses than ES (Table 1) for parameters examined. It is classified as a

"possible human carcinogen" based on thyroid tumors in rat. ToxCast true actives for CYPs (ES: hCYP2B6 & hCYP3A4; FP: hCYP2C9) correlated well with *in vivo* toxicity [7, 8]. Other true activities for ES and FP were for generalized liver enzyme activity, which could be predictive of liver toxicity observed *in vivo*.

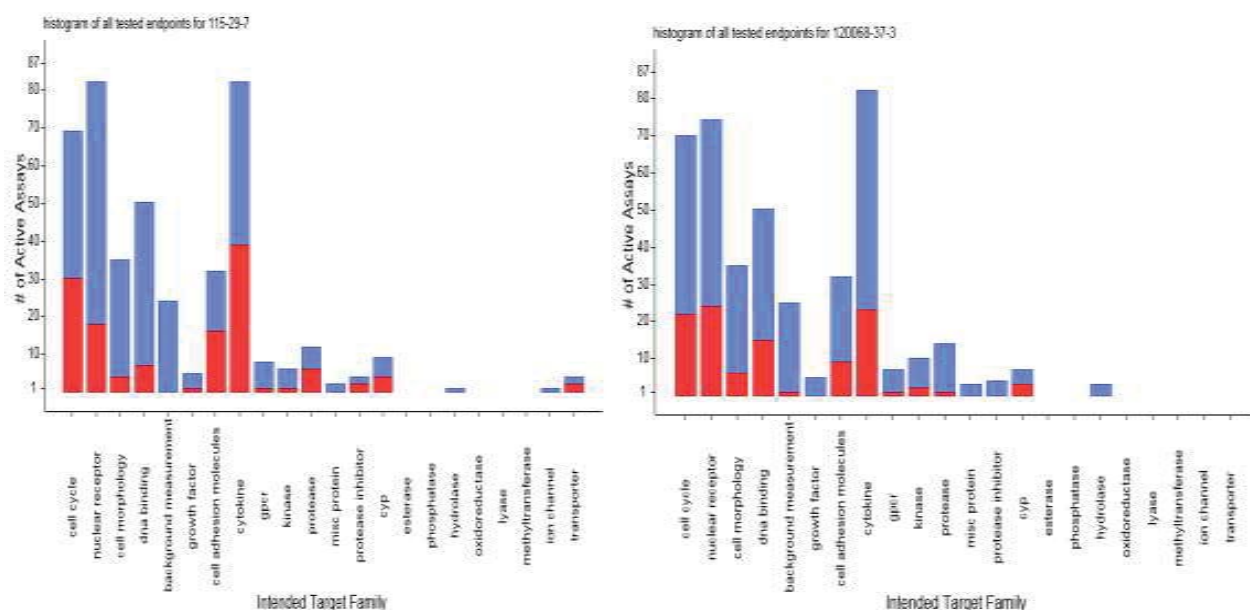
Missing were true actives for the GABA_AR ToxCast assays (0/5 total) for either compound. Since neither compound needs metabolic activation to be toxic, the inactivity could be due to a lack of adequate assay design.

ES had 2/18 total true estrogen-receptor actives and FP had 1/18. There were no true actives for the androgen receptor (0/11 total) and (especially for FP) the thyroid receptor (0/4 total). There was activity for ER, AR and TR (<http://actor.epa.gov/dashboard/2>) but only above the cytotoxicity limit). There was poor correlation between *in vivo* endocrine disruption from both chemicals and true actives with ER, AR and TR in ToxCast.

Zebrafish assays were useful for predicting developmental neurotoxicity. The OEDs reasonably correlated with NOELs observed *in vivo*: (ES: 25 fold difference; FP equivalent). With the continued rapid development of the ToxCast and zebrafish assays, they could be used to support modes of action and adverse outcome pathways for new chemicals.

Disclaimer

The opinions and conclusions expressed in this paper are those of the authors and do not necessarily represent the views or opinions of the Department of Pesticide Regulation. The authors state that their design and interpretation is not compromised by any sponsor as a condition of review and publication.



(A) Endosulfan

(A) Endosulfan

Red indicates assays considered active and blue is for inactive.

Figure 2 Proportion of active to inactive ToxCast assays by parameter measured.

Table 1 Toxicity Reported in California Department of Pesticide Regulation Risk Assessment Documents and the Open Literature^a [15-23].

Toxicity Endpoint	Endosulfan ^a	NOEL/LOEL ^b mg/kg/d	Fipronil ^a	NOEL/LOEL ^b mg/ kg/d	Fold Difference Endo:Fipronil ^c
Neurotoxicity^d	Dog: Jerky/tonic muscle contractions, convulsive movements, noise sensitivity, frightened reactions to optical stimuli, impaired reflex excitability	Dog: 0.5/2.0 ^{d 15}	Rat adult: Exaggerated startle & tail-pinch response; convulsions, tremors, ↓vision; ↓ hindleg landing splay, ↓ rearing	Rat: 0.2/2.5 ¹⁶	↓1.0
DNT^e i.p.	Rat pup: ↓ neurotransmitter binding; ↑ foot-shock fighting behavior	Rat pup: 1.0/4 ¹⁹	Rat pup: ↓ Startle response; ↓ ability to swim (unable to stay afloat, swim in straight line, or keep heads out of water)	Rat pup: 0.05/0.9 ²²	↑4.0
Thyroid Pathology/ tumors^d	No effects in any study	Not Applicable	Rat adult: follicular cell hypertrophy & hyperplasia; altered thyroid function (↓T4, ↑TSH); ↑thyroid weight; ↑follicular cell adenomas and carcinomas	Rat: 0.02/0.06 ¹⁷	Not Applicable
Liver Pathology / tumors^d	Rat adult: ↑liver weight	Rat adult: 1.92/3.95 ²⁰	Mouse adult: ↑liver weight; periadrenal vacuolation; hepato carcinomas	Mouse: 0.06/0.13 ¹⁸	↑30
Developmental^d	Rat fetus: ↓bodyweight, ↓ %live fetuses & length; ↑growth retardation, skeletal anomalies, % resorptions	Rat adult: 2.0/6.0 ²¹	Rat fetus: ↓litter survival, pup viability; delayed pinna attachment, incisor eruption, vaginal patency & preputial separation	Rat: 0.05/0.9 ²³	↑7.0

a-Lowest No-Observed-Effect-Level (NOEL) for each category.

b-NOEL/Lowest-Observed-Effect-Level (LOEL)

c-Fold difference between the lowest LOELs (where effects are observed at the lowest dose administered): Fold difference = endosulfan LOEL : fipronil LOEL

d-Endosulfan or fipronil administered orally

e- Developmental Neurotoxicity; i.p. = intraperitoneal administration of endosulfan; oral administration of fipronil

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